



The Mediterranean and Black Sea Fisheries at Risk from Overexploitation

Athanassios C. Tsikliras*, Anny Dinouli, Vasileios-Zikos Tsiros, Eleni Tsalkou

Laboratory of Ichthyology, Department of Zoology, School of Biology, Aristotle University of Thessaloniki, Thessaloniki, Greece

* atsik@bio.auth.gr



GOPEN ACCESS

Citation: Tsikliras AC, Dinouli A, Tsiros V-Z, Tsalkou E (2015) The Mediterranean and Black Sea Fisheries at Risk from Overexploitation. PLoS ONE 10(3): e0121188. doi:10.1371/journal.pone.0121188

Academic Editor: Caroline Durif, Institute of Marine Research, NORWAY

Received: September 24, 2014

Accepted: January 28, 2015

Published: March 20, 2015

Copyright: © 2015 Tsikliras et al. This is an open access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: All data are publicly available at the Food and Agricultural Organization through FishStat J software (Fishery Information, Data and Statistics Unit GFCM capture production 1970-2010. FISHSTAT J - Universal software for fishery statistical time series).

Funding: These authors have no support or funding to report.

Competing Interests: Co-author Athanassios Tsikliras is a PLOS ONE Editorial Board member. This does not alter the authors' adherence to PLOS ONE Editorial policies and criteria.

Abstract

The status of the Mediterranean and Black Sea fisheries was evaluated for the period 1970-2010 on a subarea basis, using various indicators including the temporal variability of total landings, the number of recorded stocks, the mean trophic level of the catch, the fishing-in-balance index and the catch-based method of stock classification. All indicators confirmed that the fisheries resources of the Mediterranean and Black Sea are at risk from overexploitation. The pattern of exploitation and the state of stocks differed among the western (W), central (C) and eastern (E) Mediterranean subareas and the Black Sea (BS), with the E Mediterranean and BS fisheries being in a worst shape. Indeed, in the E Mediterranean and the BS, total landings, mean trophic level of the catch and fishing-in-balance index were declining, the cumulative percentage of overexploited and collapsed stocks was higher, and the percentage of developing stocks was lower, compared to the W and C Mediterranean. Our results confirm the need for detailed and extensive stock assessments across species that will eventually lead to stocks recovering through conservation and management measures.

Introduction

Overexploitation of marine fisheries resources is either known or suspected for almost all the commercial fish stocks of the world [1]. This is the general rule, which holds globally. There are, however, a few exceptions of healthy fisheries, as a result of good management practices [2], that are justifying the rule. The issue of fisheries overexploitation at a global scale has been realized at least since 2001 [3], although there have been some earlier stock depletions that received extended news coverage, such as that of cod *Gadus morhua* in Canada in 1992 [4]. In the Mediterranean and the Black Seas (FAO Major Fishing Area 37), overexploitation has been recently shown to occur for the entire area [5,6] and for specific regions (e.g. Greek Seas [7]). Several Mediterranean stocks have been reported overexploited based on data from landings [5,7], scientific surveys [8], or stock assessments [9]. As European fisheries are no exception to the rule, much of the EU fisheries legislation actually aims to tackle with the problem of overexploitation (e.g. Council Regulations 1967/2006, 643/2007, 1224/2009, 404/2011), including the relatively recent Marine Strategy Framework Directive through its Descriptor 3 (Commission



Decision 2010/477/EU) [$\underline{10}$]. However, in parallel with global economy, management regulations are not really implemented by all parties involved, especially by those who profit for not respecting the commons; this situation may lead to ecosystem crises [$\underline{11}$].

Long-term worries about the state of stocks explain why fishery-independent stock assessments have been carried out for the past twenty years over the northern part of the Mediterranean basin [9,12], in parallel with extended monitoring of the landings that is anticipated through the EU Fisheries Data Collection Framework (Council Regulation 1999/2008). In the Mediterranean, regular scientific stock assessments are mainly carried out within two international bodies, the General Fisheries Commission for the Mediterranean (GFCM) of the Food and Agricultural Organization (FAO) and the International Commission for the Conservation of Atlantic Tunas (ICCAT), which is an inter-governmental fishery organization responsible for the conservation of tunas and tuna-like species in the Atlantic Ocean and the Mediterranean Sea. The Scientific Technical and Economic Committee for Fisheries (STECF) of the EU [12] also conducts assessments for the stocks exploited by the EU countries of the northern Mediterranean coastline. Thus, data collection, stock assessment and management of Mediterranean fisheries are gradually improving, even though there are still areas with no or inadequate fisheries information (e.g. eastern Mediterranean). Management information and guidelines resulting from stock assessments and scientific surveys, can now be delivered to the involved parties and the public at a finer geographical scale and more timely [9].

Trophodynamic indicators have been widely used in marine ecology [13] and trophic level based indicators are valuable for identifying the expansion or contraction of fisheries [14], as well as for quantifying the effects of fishing on the trophic structure of food webs [8] and the function of marine ecosystems [15], including loss of biodiversity [14]. Among the several indicators that have been used, the trophic level of the catch and the fishing-in-balance index have been extensively tested [13] and have been proven robust in tracking fishing effects [16], and assessing ecosystem changes [17]. Therefore, they are considered promising indicators for ecosystem management [13].

Mediterranean fisheries are multispecies and generally lack large monospecific stocks compared to the open ocean [18]. Consequently, the acknowledged disadvantages of stock assessments and scientific surveys, i.e., that they include a limited number of stocks compared to the total (around 25% in terms of landed biomass, less than 10% in terms of stock numbers) and are conducted at irregular basis [19], are valid for the area. Indeed, the larger proportion of landed biomass in the Mediterranean comes from data-deficient fisheries [20], i.e., stocks that have never been properly assessed, a proportion that has been estimated at around 80% [21].

The official data on fisheries catches (equivalent to landings for the purposes of this work) has been widely used to examine the state of the exploited marine resources and to test large-scale ecological hypotheses. In particular, the catch-based method of stock classification to exploitation categories has been extensively used to assess the status of fisheries globally or on ecosystem basis [5,7,22–26]. After the taxonomic, quantitative and spatial improvements through the completion of the global reconstruction of catches [27], the validity of catches will no longer be questioned. However, since there is a global debate on whether data from catches or from surveys and stock assessments are more suitable for examining the exploitation status of marine fish stocks [19], data from scientific surveys and stock assessments, when available, could be used complementary to catch data.

Earlier research on the Mediterranean and Black Sea fisheries report that 78% of the stocks are fully exploited in the Mediterranean, and 85% of the stocks are overexploited in the Black Sea [28,29]. Overexploitation was evident since the 1950s, when about 60% of the Mediterranean and Black Sea stocks were reported as being fully exploited and 40% as overexploited [30]. When the catch-based method was applied on the combined landings of Mediterranean and Black Sea,



over 80% of the Mediterranean stocks were described as fully exploited [31] and around 90% of the Black Sea stocks as collapsed, in 2004 [32]. Recently, using the same method on the same area, the cumulative percentage of collapsed and overexploited stocks exceeded 60% [5]. Based on scientific stocks assessments, 85% of the assessed stocks have been reported as being overfished compared to their maximum sustainable yield value [9]. According to the most recent work, the exploitation rate in the Mediterranean has been steadily increasing and selectivity deteriorating; thus leading to shrinkage of fish stocks [6]. Finally, a low rate decline in the mean trophic level of the catches has been reported for the Mediterranean and Black Seas [33].

In our previous work on the Mediterranean and the Black Sea fisheries [5], the stocks of the two seas were combined (as the FAO Area 37), for comparability purposes with previously published items. However, because of the different environmental and fishing properties of the two seas [28,34], their subareas and subdivisions [35,36], a further investigation of their fisheries status was necessary at a finer geographical scale, i.e., on a subarea basis.

The aim of the present work was to assess the exploitation status of the Mediterranean and Black Sea stocks for the period 1970–2010, using various indicators including the temporal variability of total landings, the number of recorded stocks, the mean trophic level of the catch, the fishing-in-balance index and the catch-based method of stock classification. These indicators were separately applied to the western, central and eastern Mediterranean Sea, thus including areas where primary data, such as fisheries independent data and stock assessment information, are lacking. Contrary to recent studies on the Mediterranean fisheries [5,9], which considered only the stocks that had been previously assessed and form a minor proportion of the commercially exploited stocks, in the present work, we considered all stocks for which landings data were available.

Materials and Methods

The annual catches of the Mediterranean and Black Sea fisheries stocks have been recorded since 1970 by the General Fisheries Commission for the Mediterranean (GFCM) [37]. The GFCM dataset differs from that of the FAO in terms of area and species aggregation. As opposed to the FAO grouping of the entire Mediterranean and the Black Seas into a single area (FAO Major Fishing Area 37), in the GFCM dataset, the Mediterranean and the Black Sea are being further divided into 4 fishing subareas (Mediterranean Sea: western, W; central, C; eastern, E; and Black Sea: BS) and 10 fishing subdivisions (Fig. 1). The GFCM data refer to the legal and reported large- and small-scale fisheries catches (excluding discarded catch, illegal, unreported, recreational, subsistence and sport fishing) expressed as live weight equivalent of landings. For the present analysis, catch statistics were extracted from the GFCM dataset, which is available as a regional package in FishStat J, for the period 1970–2010, for 2636 records (= stocks, defined here as species-area-country combinations) [37].

Two trophodynamic indicators, the mean weighted trophic level of the catch (mTLc) and the fishing-in-balance (FiB) index, that have been widely used to assess the effect of fisheries on ecosystems [38], were explored. The mean weighted trophic level of the catch (mTLc) for each year (k), was calculated annually per area according to the formula [33]:

$$mTLc = \frac{\sum_{i=1}^{m} (TL_i \times Y_{ik})}{\sum_{i=1}^{m} (Y_{ik})}$$
 (1)

where Y_i refers to the catches of a species (or group of species) i, and TL is the corresponding trophic level. The trophic levels of each species were taken from FishBase [39] and SeaLifeBase [40].



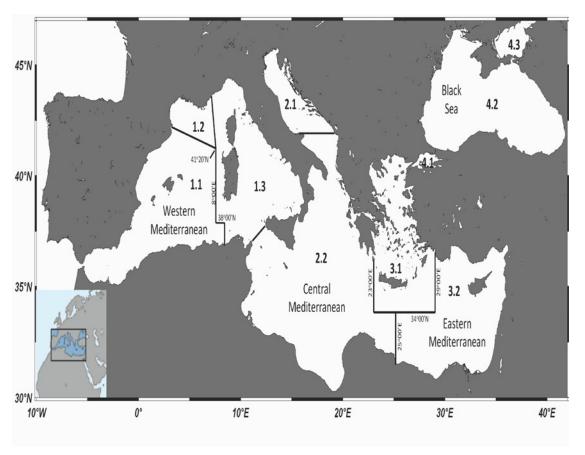


Fig 1. A map of the Mediterranean and the Black Sea (FAO Major Fishing Area 37) subareas (western Mediterranean, central Mediterranean, eastern Mediterranean, Black Sea) and their subdivisions (1.1: Balearic; 1.2: Gulf of Lions; 1.3: Sardinia; 2.1: Adriatic Sea; 2.2: Ionian Sea; 3.1: Aegean Sea; 3.2: Levantine Sea; 4.1: Marmara Sea; 4.2: Black Sea, proper; 4.3: Azov Sea) according to the GFCM division. Underlying map was created using Ocean Data View Software.

The fishing-in-balance index (FiB) of the catch for each year (k) was calculated annually per area as [41]:

$$FiB_{k} = log_{10} \left[Y_{k} \left(\frac{1}{TE} \right)^{mTLC_{k}} \right] - log_{10} \left[Y_{0} \left(\frac{1}{TE} \right)^{mTLC_{0}} \right]$$
 (2)

where Y refers to the total catches in year k, mTLc is the mean weighted trophic level of the catch, TE is the mean energy-transfer efficiency between trophic levels that is assumed to be 0.1 [14,41], and 0 refers to the first year in a time-series that is used as a baseline. In the present dataset, the first year of the time-series, 1970, was set as a baseline. FiB equals to 0 for the first year of the series and remains rather stable when trophic level and catches change in opposite directions. Increasing FiB values indicate geographic or bathymetric expansion of fisheries, while decreasing FiB values indicate contraction [41].

The exploitation status of fisheries (1970–2010) was classified into one of the following five categories, according to the catch-based stock classification method [$\underline{26,30}$]: developing, fully exploited, overexploited, collapsed, and recovering of a previously collapsed stock ($\underline{Table\ 1}$). The "undeveloped" category was incorporated in "developing". The classification is based on the relationship between the catches (C_Y) of a given year (Y_C) compared to the year (Y_{Cmax}) of



Table 1. The five categories (developing, fully exploited, overexploited, collapsed and recovering of a previously collapsed stock) of stock exploitation according to the catch-based method, which is based on the relationship between the catches (C_Y) of a given year (Y_C) compared to the year (Y_{Cmax}) of historical maximum catch (C_{MAX}) .

Exploitation status	Criteria
Developing	$Y_C < Y_{Cmax}$ and $0.1C_{MAX} < C_Y < 0.5C_{MAX}$
Fully exploited	$C_Y > 0.5C_{MAX}$
Overexploited	$Y_C > Y_{Cmax}$ and $0.1C_{MAX} < C_Y < 0.5C_{MAX}$
Collapsed	$Y_C > Y_{Cmax}$ and $C_Y < 0.1 C_{MAX}$
Recovering of a previously collapsed stock	$Y_C > Y_{Cmax}$ and $0.1C_{MAX} < C_Y$

historical maximum catch (C_{MAX}) (<u>Table 1</u>). Stocks with maximum catches appearing in the most recent year of the time series were designated as "developing" [26].

The method was applied on an annual basis to the stocks of each subarea (n = 2207 for the last year of the time series, 2010) and to the stocks of each subdivision (n = 1707 for the last year of the time series, 2010). The method was then separately applied to those stocks whose catches are available for 40 consecutive years (n = 802 for the last year of the time series, 2010). The latter analysis was included because it has been shown that the addition of new records (either because of disaggregation of higher taxonomic groups or because of exploitation of new species/stocks) in the dataset, may introduce biases in favour of developing stocks [5,7]. The number of records varies among analyses because several records were only available for nonconsecutive years, some were collected from unknown or unspecified subareas and some from the entire Mediterranean Sea, such as the highly migratory bluefin tuna *Thunnus thynnus* [37].

Finally, the results and recommendations of the stock assessments routinely performed by the STECF and GFCM were retrieved from the published reports [42-46] that refer only to the Mediterranean Sea. Overall, 189 records were accessed (STECF: 18 species, 78 stocks; GFCM: 25 species, 111 stocks) from which information on current fishing mortality (F) and fishing mortality level able to provide maximum sustainable yield (F_{MSY}) was extracted (for details on stock assessment methodology, assumptions, limitations and approximations see [9]). MSY is defined as the highest yield (= catch) that can be taken from a stock under existing environmental conditions, and F_{MSY} is the fishing mortality that allows stock exploitation at MSY [47,48]. The F/F_{MSY} ratio, which indicates sustainable exploitation when lower or close to 1 and unsustainable when diverging to high values [47,49], was then compared with the ratio Y_{2010}/Y_{MAX} (where Y_{2010} is the latest catch and Y_{MAX} is the highest catch in the catch time series), which shows the exploitation status of the corresponding stocks according to the catch-based method. A paired match was possible for 97 stocks, either because of multiple assessments of the same stock (when multiple assessments were available for a stock, the most recent one was used) or because of unavailability of Y_{2010}/Y_{MAX} records.

Results

Number of records

The number of recorded fish stocks in the W and C Mediterranean increased linearly ($r^2>0.88$ in both areas) from around 200 in 1970 to around 600 in 2010 (Fig. 2). A linear increase was also apparent for the E Mediterranean ($r^2=0.90$) and the Black Sea ($r^2=0.89$), but with fewer records (from around 150 in 1970 to around 350 in 2010 in E Mediterranean and from around 100 in 1970 to around 200 in 2010 in the Black Sea) (Fig. 2). The rate of increase for the entire study period was around 80 records per decade (i.e., the catches of 80 new stocks were added



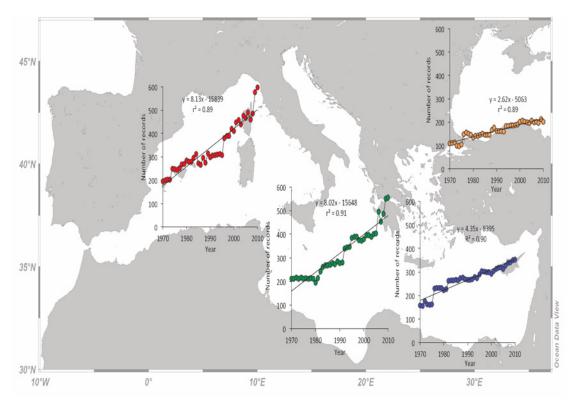


Fig 2. The number of stocks recorded per year for the western (red dots), central (green dots), eastern (blue dots) Mediterranean subareas and the Black Sea (orange dots) for the period 1970 to 2010.

or separately recorded) for the W and C Mediterranean and lower for the E Mediterranean (44 records per decade) and the Black Sea (26 records per decade).

Total catches

Total catches in the W Mediterranean fluctuated around 350 000 t with 304 000 t landed in 2010 (minimum catch was 295 000 t in 2007 and maximum catch was 433 467 t in 2006) and, overall, do not show any increasing or declining trend (Fig. 3). In C Mediterranean, total catches increased rapidly from 228 400 t in 1970 to 466 218 t in 1984, then declined and fluctuate around 350 000 t for the last two decades (the catch of 2010 was 342 083 t) (Fig. 3). A similar trend was observed for the E Mediterranean but total catches were lower, the increase was smoother and lasted longer (from 66 318 t in 1970 to 295 870 t in 1994). Since 1994, total catches in E Mediterranean fluctuate around 200 000 t, with catches of 216 804 t reported in 2010 (Fig. 3). The Black Sea total catches peaked between 1980 and 1988, with values exceeding 800 000 t per year. For the periods before (1970–1979) and after that peak (1989 onwards), total catches in the Black Sea fluctuated around 450 000 t, with 501 241 t reported in 2010 (Fig. 3).

Mean trophic level of the catch (mTLc) and fishing-in-balance (FiB) index

There was a clear and steady decline of mTLc in the W Mediterranean for the period 1970–2004, from around 3.30 (1970–1985) to around 3.20 (1986–2004), at a rate of 0.03 per decade. Since 2005, mTLc increased again to around 3.35, with an irregular high peak in 2007 (Fig. 4). In the C Mediterranean, mTLc fluctuated around 3.20 for the entire period, with a five year



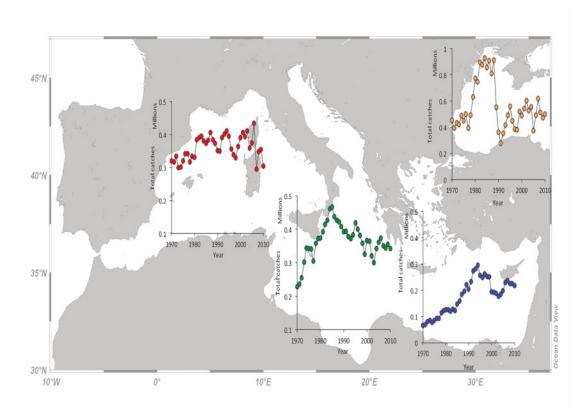


Fig 3. The combined marine catches (in metric tons) of fishes, crustacean and cephalopods per year for the western (red dots), central (green dots), eastern (blue dots) Mediterranean subareas and the Black Sea (orange dots) for the period 1970 to 2010.

exception (1999–2004) of lower mTLc values at around 3.10 (Fig. 4). The declining trend of mTLc in the E Mediterranean is clear and consistent throughout the study period. The rate of mTLc decline is 0.03 per decade, from 3.38 in 1970 to 3.25 in 2010 (Fig. 4). A quicker declining trend of the mTLc is also observed in the Black Sea since 1988, following an initial increase from 1970 to 1987. The rate of mTLc decline (1988–2010) is 0.09 per decade, from 3.42 in 1988 to 3.22 in 2010 (Fig. 4).

Given the rather stable catches, the FiB index in the W Mediterranean followed a similar pattern with the mTLc, with elevated values since 2005 (Fig. 5). Similarly, in the C Mediterranean, where mTLc was rather stable, the fluctuation of FiB was determined by the catches. It was initially increasing from 1970 to 1985, then declined until the mid 2000s before rising up again (Fig. 5). In the E Mediterranean, FiB rapidly increased from 1970 to its highest value in 1994 (FiB = 0.59 in 1994) and then declined to around 0.38 in 2010 (Fig. 5). In the Black Sea, an initial increase of FiB from 1970 to its highest value in 1988 (FiB = 0.43 in 1988) was followed by a rapid decline and stabilization to values around 0 (Fig. 5).

Exploitation status according to the catch-based method

Based on the year and quantity of the historically maximum catch, which varied among the stocks included in the analysis, in 2010, the cumulative percentage of overexploited and collapsed stocks was 60% for the Black Sea, 56% for the W, 48% for the E, and 44% for the C Mediterranean (Fig. 6). In contrast, the cumulative percentage of developing and recovering stocks



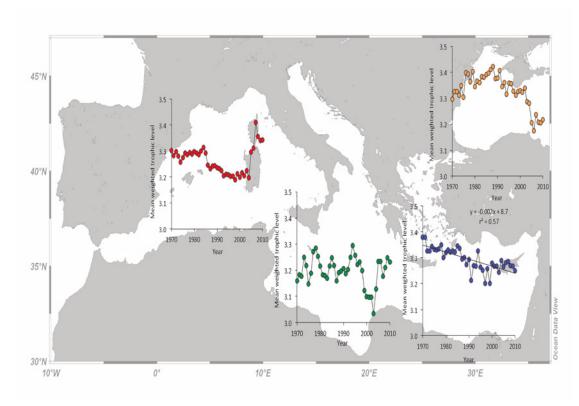


Fig 4. The mean trophic level of the catch (mTLc) per year for the western (red dots), central (green dots), eastern (blue dots) Mediterranean subareas and the Black Sea (orange dots) for the period 1970 to 2010.

was higher in the C Mediterranean (28%) and the Black Sea (26%) compared to the W (24%) and E Mediterranean (21%) (Fig. 6).

The percentage of developing stocks declined linearly with time across subareas, while, at the same time, the cumulative percentage of overexploited and collapsed stocks increased linearly (Fig. 7). The higher decline in developing stocks was observed in the E Mediterranean and the lower decline in the W and C parts of the sea. Indeed, the rate of decline in developing stocks was 21, 20, 54 stocks per decade in the W, C, E Mediterranean, respectively, and 30 stocks per decade in the Black Sea. The higher increase in the cumulative percentage of overexploited and collapsed stocks was observed in the W and C Mediterranean and the lower increase in the Black Sea (Fig. 7). The rate of increase in overexploited and collapsed combined stocks was 86, 76, 62 stocks per decade for the W, C, E Mediterranean, respectively, and 46 stocks per decade for the Black Sea.

When only stocks with more than 40 consecutive records were considered, the cumulative percentage of overexploited and collapsed stocks in 2010 increased across subareas. It was 72% for the Black Sea, 65% for the W Mediterranean, 62% for the C Mediterranean and 61% for the E Mediterranean (Fig. 8). The cumulative percentage of developing and recovering stocks was rather stable among subareas: 22% for the Black Sea, 18% for the W Mediterranean, 19% for the C Mediterranean and 18% for the E Mediterranean (Fig. 8). Developing stocks were declining at a rate of 25, 36, 44 stocks per decade for the W, C, E Mediterranean, respectively, and 21 stocks per decade for the Black Sea. Overexploited and collapsed stocks were increasing at a rate of 44, 33, 38 stocks per decade for the W, C, E Mediterranean, respectively, and 13 stocks per decade for the Black Sea.



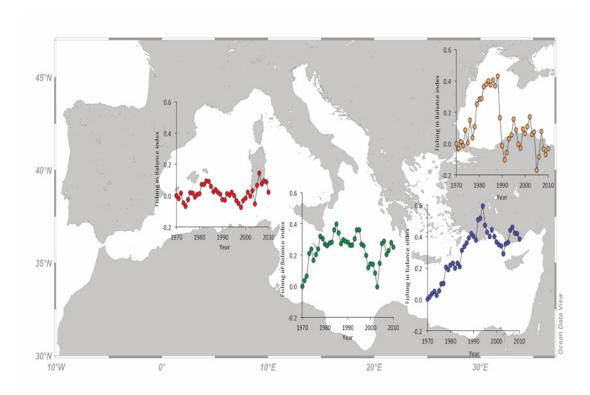


Fig 5. The Fishing-in-Balance (FiB) index per year for the western (red dots), central (green dots), eastern (blue dots) Mediterranean subareas and the Black Sea (orange dots) for the period 1970 to 2010.

The exploitation pattern differed among the subdivisions of each subarea of the Mediterranean and the Black Sea (Table 2). In 2010, the cumulative percentage of overexploited and collapsed stocks exceeded 50% in all fishing subdivisions, except subdivisions 2.1 (Adriatic) and 3.2 (Levantine), which presented the lowest cumulative percentage of overexploited and collapsed stocks (Table 2). The degree of exploitation also varied among subdivisions. Subdivisions 1.2 (Lions Gulf), 3.1 (Aegean Sea), 4.1 (Marmara Sea) and 4.2 (main Black Sea) were the most heavily exploited ones with the cumulative percentage of overexploited and collapsed stocks exceeding 55% (Table 2). The highest percentage of developing fisheries stocks were recorded in subdivision 2.1 (Adriatic) and the highest percentage of recovering stocks in subdivisions 4.2 (main Black Sea), 4.1 (Marmara) and 3.2 (Levantine) (Table 2).

Comparison of the catch-based method with stock assessment data

The relationship between the F/F_{MSY} ratio and the Y_{2010}/Y_{MAX} ratio was expressed by an inverse power function (Fig. 9), which shows that high fishing mortality rates may lead to low yields and stock collapses. According to the theoretical plot of this relationship (Fig. 10, sensu [50,51]), out of the 97 stocks from the entire Mediterranean Sea that were analyzed (Fig. 9), 44 stocks (45%) were fully exploited, 48 (49%) were overexploited, and 5 stocks (6%) were collapsed. No undeveloped/developing stocks were recorded (data not shown).

Discussion

The results of the present work confirm, through the use of various indicators, that the fisheries resources of the Mediterranean and Black Sea are at risk from overexploitation. Similar



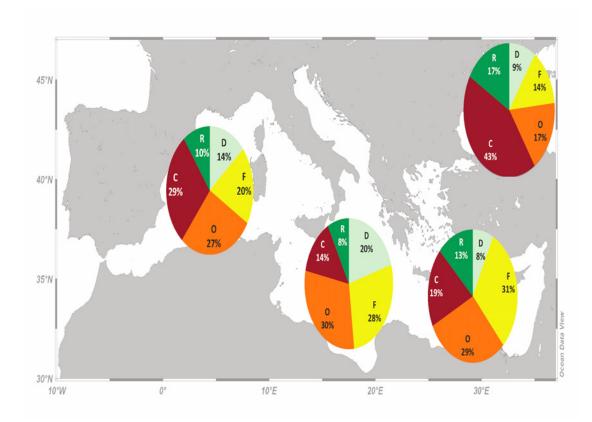


Fig 6. The status in 2010 of fisheries resources according to the catch-based method, for the western, central, eastern Mediterranean fishing subarea and the Black Sea for the period 1970 to 2010 (D: developing; F: fully exploited; O: overexploited; C: collapsed; R: recovering).

conclusions on the overexploitation of these areas have been recently reported based on various datasets and approaches [5,6,19,52]. According to all indicators that were explored here, as summarized in Table 3, the pattern of exploitation and the state of stocks differs among subareas. All indicators confirmed the worst shape of the E Mediterranean and Black Sea fisheries compared to the W and C Mediterranean ones, none of which is underexploited. Indeed, in the E Mediterranean and the Black Sea, total catches, mTLc and FiB are declining, the cumulative percentage of overexploited and collapsed stocks exceeded 50% and developing stocks were less than 10%. In contrast, in the W Mediterranean, total catches are rather stable, mTLc and FiB are increasing following a long decline, and developing stocks exceed 10%. Finally, in the C Mediterranean, mTLc and FiB are also increasing following a long decline, the cumulative percentage of overexploited and collapsed stocks do not exceed 50% and developing stocks are 20%.

The number of records increased in all subareas as a result of the disaggregation and the finer taxonomic resolution in catch records, partly because of the EU Fisheries Data Collection Framework that is applied to the European countries of the northern Mediterranean coastline. The higher number of records in the W and C Mediterranean compared to the E one (Fig. 2) agrees with the general biodiversity pattern in the Mediterranean Sea and reflects the number of species, which declines from northwestern to southeastern regions following the primary production gradient [35]. Specifically, the number of fish species that form the majority of catches decreases from western to eastern regions and is generally higher along the Mediterranean coastline [35].



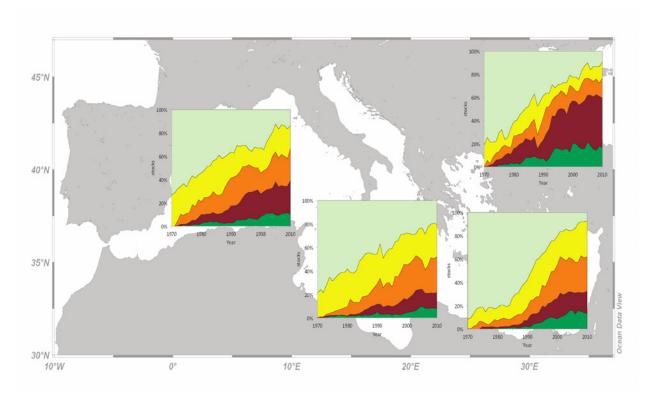


Fig 7. The trend for all stocks according to the catch-based method, for the western, central, eastern Mediterranean fishing subarea and the Black Sea for the period 1970 to 2010 (light green: developing; yellow: fully exploited; orange: overexploited; brown: collapsed; dark green: recovering).

Given that global fishing effort remains the same or increases with time, as it also does in the Mediterranean Sea [53], a decline in total catches, which is actually a decline in catch per unit of effort (CPUE), can be considered as a suspicion or indication of overfishing. There are a few well-documented examples where CPUE successfully tracks the stock size [54] but this is not always the case. In cases where fishing mortality follows predictable dynamics over time, catch data can be used to reconstruct biomass and potentially used for conducting data-poor assessments [55]. Declines in CPUE have been reported for the vast majority of FAO areas, including the Mediterranean Sea, where a drastic CPUE decline is apparent since the late 1990s [1]. The more or less stable total catches of the W Mediterranean, but with a declining trend during the last decade, is probably the result of exploiting new species, as revealed by the number of added records per year (Fig. 2), and to improving management practices as revealed by the larger number of stocks that are regularly assessed [9]. The declining trend of C Mediterranean total catches after mid 1980s and of the E Mediterranean ones after mid 1990s (Fig. 3), is a clear effect of excessive fishing mortality applied on these stocks [42] and the alteration and simplification of the food web structure over time as a result of fishing [35,56]. Overfishing and its side effects on ecosystem have been blamed for the collapse of Black Sea fisheries in early 1990s and the decline in total catches [28]. Some catch fluctuations, however, especially those of the small and medium pelagic fishes that are highly represented in the catches, could be the result of climatic effects [57] or modifications in the planktonic production [58].

The trends of mTLc were declining for all subareas until about 2005, when they started to increase again for the W and C Mediterranean (Fig. 4). The abrupt increase in the mTLc in the W Mediterranean after 2005 (Fig. 4) may be attributed to either a decline in the catches of low



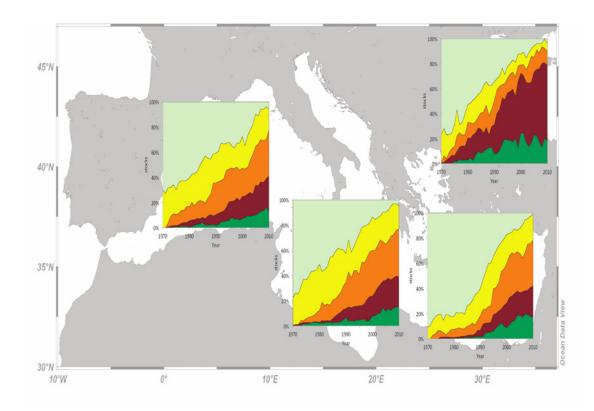


Fig 8. The trend considering only the stocks with 40 consecutive records according to the catch-based method, for the western, central, eastern Mediterranean fishing subarea and the Black Sea for the period 1970 to 2010 (light green: developing; yellow: fully exploited; orange: overexploited; brown: collapsed; dark green: recovering).

Table 2. The status in 2010 of fisheries resources according to the catch-based method per GFCM fishing subdivision (area codes as they appear on Fig. 1) of the Mediterranean and Black Sea.

			Overexploited Number (%)	Collapsed Number (%)	Recovering Number (%)	Total stocks
Exploitation Area	Developing Number (%)	Fully exploited Number (%)				
Balearic (1.1)	25 (11)	67 (30)	69 (31)	44 (20)	18 (8)	223
Gulf of Lions (1.2)	26 (15)	19 (11)	34 (20)	70 (42)	20 (12)	169
Sardinia (1.3)	35 (17)	37 (18)	64 (32)	45 (22)	22 (11)	203
Adriatic (2.1)	52 (26)	53 (27)	48 (24)	24 (12)	22 (11)	199
Ionian (2.2)	63 (18)	94 (26)	125 (35)	54 (15)	21 (6)	357
Aegean (3.1)	4 (3)	32 (25)	52 (40)	28 (22)	13 (10)	129
Levantine (3.2)	24 (11)	69 (30)	59 (26)	40 (18)	33 (15)	225
Black Sea						
Marmara (4.1)	0 (0)	4 (8)	12 (25)	24 (50)	8 (17)	48
main Black Sea (4.2)	16 (14)	13 (11)	17 (14)	49 (42)	23 (19)	118
Azov Sea (4.3)	5 (14)	11 (31)	7 (19)	11 (31)	2 (5)	36
Total						1707

doi:10.1371/journal.pone.0121188.t002



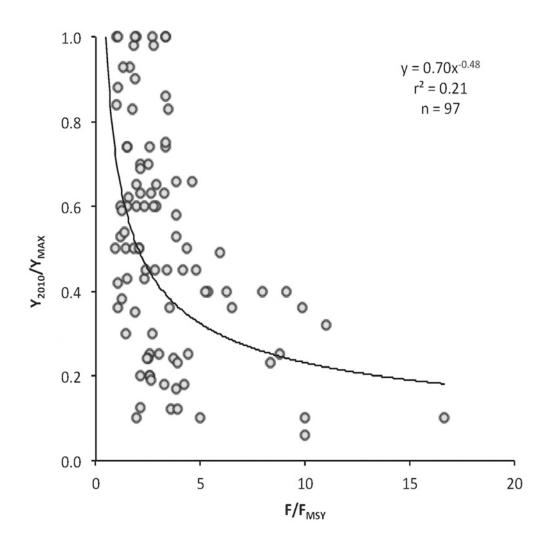


Fig 9. The relationship of the ratio of the current fishing mortality (F) and mortality at MSY (F_{MSY}) with the ratio of the latest catch in 2010 (Y_{2010}) to historically maximum catch (Y_{MAX}) for 97 stocks for which both ratios were available (for explanation see maintext).

trophic level species, or an increase in the catches of high trophic level species. In any case, this is a clear situation of fishing-up the marine food webs [8]. This mTLc increase, together with a simultaneous decline in catches, produced a rather stable FiB trend with a slight increasing tendency during the last decade (Fig. 5). Since 2007, it seems that at least one ecosystem of the W Mediterranean (Gulf of Lions) has shifted to a different regime, which is characterized by low sardine and anchovy catches [58]. According to the same work [58], the landings of sardine and anchovy declined by 50% and 70%, respectively, which confirms the fishing-up effect of the last lustrum. Similarly, a decline in top predators and large sized fish, as well as an increase of lower trophic level species has been reported for another ecosystem of the W Mediterranean, the Catalan Sea [56]. The higher exploitation rate of demersal and large pelagic stocks in the Adriatic Sea since 1990 has lead to mTLc decline until about 2005, when mTLc started to increase again (Fig. 4). The exact same pattern was apparent in FiB (Fig. 5), which is a function of the mTLc, when catches are stable or follow the same trend. Similarly to the W Mediterranean, a long-term decline in large-sized, high trophic level species has been reported to occur during



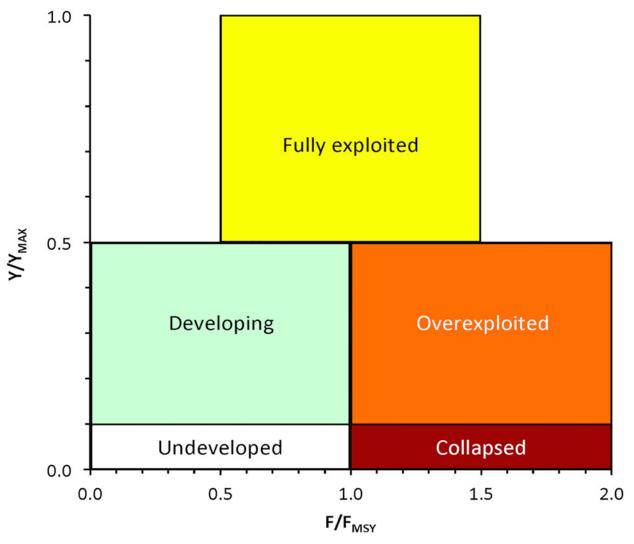


Fig 10. Theoretical plot of the stock status definitions (as they appear on Table 1, but excluding "recovering" for simplicity purposes) based on the relationship between the ratio of the latest catch (Y) to historically maximum catch (Y_{MAX}) and their relationship with the ratio of current fishing mortality (F) to mortality at MSY (F_{MSY}).

Table 3. Overview of all indicators (TC: total catch; CBM: catch-based method; mTLc: mean trophic level of the catch; FiB: fishing in balance index) used in the present work for evaluating the status of Mediterranean and Black Sea fisheries per fishing subarea, indicating their condition in 2010.

Indicator Area	тс	СВМ	mTLc	FiB	Overall
Mediterranean Sea					
Western	Average	Bad	Average	Average	Average
Central	Bad	Average	Average	Average	Average
Eastern	Bad	Bad	Bad	Bad	Bad
Black Sea					
Black Sea	Bad	Bad	Bad	Bad	Bad

Good condition of a time-series refers to increasing, bad to declining and average to stable trends.

doi:10.1371/journal.pone.0121188.t003



the same period for the Adriatic Sea [59], together with an increase in species positioned lower in the food web [60]. After 2005, the situation seems reversed, a pattern that was apparent in the FiB values as well. In the E Mediterranean Sea, the decline in mTLc (Fig. 4), albeit at low rates compared to other areas, was continuous. Because of the low rate of mTLc decrease, FiB followed the fluctuation of catches [38]. A decline in mTLc for species with trophic level greater than 3.75 has been shown to occur for the Greek Seas [61], whereas stable mTLc values for all species have been recently reported for the same area [7]. Based on the latter, the decline in mTLc is more likely to occur because the Aegean Sea is being exploited both by the Greek and Turkish fishing fleets. Indeed, the trawling and dredging disturbance as well as the exploitation pattern is very high along the Aegean coastline and lower in the Levantine Sea [62], where the stocks are generally in a better shape (Table 2). The collapse of the Black Sea fisheries in 1990 and the decline in the mTLc since then, is apparent in the FiB values that declined accordingly (Figs. 4 and 5). Indeed, a decrease in abundance of high trophic level species and an increase of low trophic level species has been reported to occur after 1990 for the Black Sea [63]. One of the reasons for the collapse of fisheries in the Black Sea was the invasion of the ctenophore Mnemiopsis leidyi, which preys upon the eggs and larvae of pelagic fish and may cause population declines [64].

Regarding the status of exploitation according to the catch-based method (Figs. 6 and 7), a few points can be considered encouraging compared to the current global pattern. Firstly, the percentage of developing stocks, and hence developing fisheries, is generally higher in all Mediterranean subareas and, secondly, the cumulative percentage of overexploited and collapsed ones is lower, when compared to an analysis on global fisheries [65] and the northeast Atlantic [51]. Developing stocks indicate the exploration of new fisheries and exploitation of new stocks or increasing stock biomass that is reflected upon catches, whereas recovering stocks indicate good or improving management practices applied on fisheries that had been previously depleted or collapsed [26]; but see also [66]. In the Adriatic Sea (C Mediterranean) the vast majority of demersal and large pelagic stocks had been classified as overfished since 2000, with the small and medium pelagic ones being in a better state [60]. In the Aegean Sea (E Mediterranean), about 65% of the Greek stocks were characterised as overexploited and 32% as fully exploited in 2007 [7], whereas the 85% of the Black Sea stocks are overexploited [29].

The use of stocks with consecutive records may limit the number of stocks included in the analysis but shows the fate of the catches of the same community through time and it is far less biased by the number of new records added each year [5]. When new species are added separately in the analysis (after disaggregation), they are by definition developing. Therefore, the percentage of developing stocks is being artificially increased and, at the same time, the percentage of the remaining categories proportionally decreased [7]. This bias is eliminated only when stocks with non-consecutive records are excluded and for that reason we re-analyzed the dataset including only stocks with consecutive records (Fig. 8). Once this effect was removed, the cumulative percentage of overexploited and collapsed stocks increased across subareas and that of the developing ones decreased (Fig. 8). The effect was more intense in W and C Mediterranean, where the number of stocks added each year was higher. This clearly indicates that the addition of new records may mask the true condition of stocks by increasing the percentage of developing ones [5]. The rate of increase of overexploited and collapsed consecutive stocks in the Black Sea was lower compared to other subareas because the actual number of stocks with consecutive records was very low (n = 78) compared to the non-consecutive ones (n = 283).

Finally, when the data from stock assessments was assembled and compared to the catch-based method (Fig. 9), the pattern of the relationship between the F/F_{MSY} and the Y_{2010}/Y_{MAX} was similar to the F/F_{MSY} against B/B_{MSY} that has been recently published [50]. A comparison



of Y_{2010}/Y_{MAX} with B/B_{MSY} would have been preferable because catch-based status predictions refer to biomass and not fishing mortality. However, B/B_{MSY} values are not as commonly used in the Mediterranean stock assessments and are more biased compared to F/F_{MSY} estimates. The results of this relationship (according to Fig. 10), i.e., that the majority of recently assessed stocks are either fully exploited or overexploited, imply that assessments are applied to stocks for which there is suspicion or indication for overexploitation. When not routinely performed, as it is the case for most Mediterranean species, stock assessments can be considered as snapshots of the stock state that may not necessarily reflect its true condition. The catch-based method may have some disadvantages related to the quality of data [50], but it provides a time series of the condition of all stocks in one area and can be used as a first indicator of ecosystem health or its response to fishing [26]. It should be noted here that, in general, trophodynamic indicators are considered rather conservative indicators of ecosystem health because they tend to respond slowly to structural ecosystem changes [38].

The findings of the present work, through the use of various indicators, clearly show that the Mediterranean and Black Sea fisheries resources are at risk from overexploitation and that the degree of exploitation varies among subareas. They also sound the alarm bell for the remaining stocks and warn that cases of mismanagement should be abandoned [67]. Detailed and extensive stock assessments are required in relation to stock status and fishing mortality reference points, that will eventually lead to conservation policy through management measures.

Acknowledgments

The authors would like to thank the Academic Editor of PLOS ONE Caroline Durif and two anonymous reviewers for their very helpful comments that really improved our work.

Author Contributions

Conceived and designed the experiments: ACT. Performed the experiments: ACT AD VZT ET. Analyzed the data: ACT AD VZT ET. Wrote the paper: ACT.

References

- Watson RA, Cheung WWL, Anticamara JA, Sumaila RU, Zeller D, Pauly D. Global marine yield halved as fishing intensity redoubles. Fish Fish 2013; 14: 493–503
- Hilborn R. Reinterpreting the state of fisheries and their management. Ecosystems 2007; 10: 1362–1369
- Watson R, Pauly D. Systematic distortions in world fisheries catch trends. Nature 2001; 414: 534–536 PMID: 11734851
- Hutchings JA, Myers RA. What can be learned from the collapse of a renewable resource? Atlantic Cod, Gadus morhua, of Newfoundland and Labrador. Can J Fish Aquat Sci 1994; 51: 2126–2146
- Tsikliras AC, Dinouli A, Tsalkou E. Exploitation trends of the Mediterranean and Black Sea fisheries. Acta Adriat 2013; 54: 273–282
- Vasilakopoulos P, Maravelias CD, Tserpes G. The alarming decline of Mediterranean fish stocks. Curr Biol 2014; 24: 1643–1648 doi: 10.1016/j.cub.2014.05.070 PMID: 25017210
- Tsikliras AC, Tsiros V-Z, Stergiou KI. Assessing the state of Greek marine fisheries resources. Fish Manag Ecol 2013; 20: 34–41
- 8. Stergiou KI, Tsikliras AC. Fishing-down, fishing-through and fishing-up: fundamental process versus technical details. Mar Ecol Progr Ser 2011; 441: 295–301
- Colloca F, Cardinale M, Maynou F, Giannoulaki M, Scarcella G, Jenko K, et al. Rebuilding Mediterranean fisheries: a new paradigm for ecological sustainability. Fish Fish 2013; 14: 89–109
- Borja A, Elliott M, Andersen JH, Cardoso AC, Carstensen J, Ferreira JG, et al. Good Environmental Status of marine ecosystems: What is it and how do we know when we have attained it? Mar Poll Bull 2013; 76: 16–27



- Tsikliras AC, Sumaila UR, Stergiou KI. Parallels in economic and ecosystem crises. Ethics Sci Environ Polit 2013; 13: 23–25
- Lleonart J. Review of the state of Mediterranean and Black Sea fisheries resources. Options Medit, Ser. B 2008; 62: 57–69
- Libralato S, Pranovi F, Stergiou KI, Link JS. Trophodynamics in marine ecology: 70 years after Lindeman. Mar Ecol Prog Ser 2014; 512: 1–7
- Kleisner K, Mansour H, Pauly D. Region-based MTI: resolving geographic expansion in the Marine Trophic Index. Mar Ecol Prog Ser 2014; 512: 185–199
- Shannon L, Coll M, Bundy A, Gascuel D, Heymans JJ, Kleisner K, et al. Trophic level-based indicators to track fishing impacts across marine ecosystems. Mar Ecol Prog Ser 2014; 512: 115–140
- Pranovi F, Link J, Fu C, Cook AM, Liu H, Gaichas S, et al. Trophic level determinants of biomass accumulation in marine ecosystems. Mar Ecol Prog Ser 2012; 459: 185–201
- Pranovi F, Libralato S, Zucchetta M, Link JS. Biomass accumulation across trophic levels: analysis of landings for the Mediterranean Sea. Mar Ecol Prog Ser 2014; 512: 201–216
- Lleonart J, Maynou F. Fish stock assessments in the Mediterranean: state of the art. Sci Mar 2003; 67 (S1): 37–49
- Pauly D, Hilborn R, Branch TA. Does catch reflect abundance? Nature 2013; 494: 303–306 doi: 10.38/494303a PMID: 23426308
- Pilling GM, Apostolaki P, Failler P, Floros C, Large PA, Morales-Nin B, et al. Assessment and management of data-poor fisheries. In: Payne A, Cotter J, Potter T, editors. Advances in Fisheries science: 50 years on from Beverton and Holt. Blackwell Publishing; 2008. pp. 280–305.
- Le Quesne W, Brown M, De Oliveira J, Casey J, O'Brien C. Data-deficient fisheries in EU waters. Directorate-General for Internal Policies, Policy Department B: Structural and Cohesion Policies. Fisheries. IP/B/PECH/IC/2012–118. 2013. pp.7
- Worm B, Barbier EB, Beaumont N, Duffy JE, Folke C, Halpern BS, et al. Impacts of biodiversity loss on ocean ecosystem services. Science 2006; 314: 787–790 PMID: 17082450
- 23. Pauly D. Global fisheries: a brief review. J Biol Res 2008; 9: 3–9
- Zeller D, Cheung W, Close D, Pauly D. Trends in global marine fisheries-a critical view. In: Wrammer P, Ackefors H, Cullberg M, editors. Fisheries, Trade and Development. Stockholm: Royal Swedish Academy of Agriculture and Forestry; 2008. pp. 55–77
- Froese R, Kesner-Reyes K. Out of new stocks in 2020: a comment on "Not all fisheries will be collapsed in 2048". Mar Pol 2009: 33: 180–181
- Kleisner K, Froese R, Zeller D, Pauly D. Using global catch data for inferences on the world's marine fisheries. Fish Fish 2013; 14: 293–311
- Zeller D, Pauly D. Reconstruction of marine fisheries catches for key countries and regions (1950–2005).
 Fisheries Centre Research Reports 2007; 15(2): 1–163
- Daskalov GM. Overfishing drives a trophic cascade in the Black Sea. Mar Ecol Progr Ser 2002; 225: 53–63
- 29. Sherman K, Adams S. Sustainable Development of the World's Large Marine Ecosystems during Climate Change: A commemorative volume to advance sustainable development on the occasion of the presentation of the 2010 Goteborg Award. Switzerland: IUCN; 2010
- Froese R, Kesner-Reyes K. Impact of fishing on the abundance of marine species. ICES Council Meeting Report CM 2002; 12/L: 1–15
- 31. Aquarone MC, Adams S, Mifsud P. Mediterranean Sea LME. In: Sherman K, Hempel G, editors. The UNEP Large Marine Ecosystem Report: A perspective on changing conditions in LMEs of the world's Regional Seas. UNEP Regional Seas, Report and Studies No. 182. UN Environment Programme; 2008. pp. 189–200.
- 32. Heileman S, Parr W, Volovik G. Black Sea LME. In: Sherman K, Hempel G, editors. The UNEP Large Marine Ecosystem Report: A perspective on changing conditions in LMEs of the world's Regional Seas. UNEP Regional Seas, Report and Studies No. 182. UN Environment Programme; 2008. pp 203–218
- Pauly D, Christensen V, Dalsgaard J, Froese R, Torres F Jr. Fishing down marine food webs. Science 1998; 279: 860–863 PMID: 9452385
- 34. FAO. Review of the state of world fishery resources: Marine fisheries. Rome: FAO; 1997.
- 35. Coll M, Piroddi C, Steenbeek J, Kaschner K, Ben Rais Lasram F, Aguzzi J, et al. The biodiversity of the Mediterranean Sea: estimates, patterns, and threats. PLoS ONE 2010; 5(8): e11842 doi: 10.1371/journal.pone.0011842 PMID: 20689844



- **36.** Tsikliras AC, Antonopoulou E, Stergiou KI. Spawning period of Mediterranean marine fishes. Rev Fish Biol Fisher 2010; 20: 499–538
- FAO. Fishery Information, Data and Statistics Unit GFCM capture production 1970–2010. FISHSTAT J

 —Universal software for fishery statistical time series; 2013
- Cury PM, Shannon LJ, Roux J-P, Daskalov GM, Jarre A, Moloney CL, et al. Trophodynamic indicators for an ecosystem approach to fisheries. ICES J Mar Sci 2005; 62: 430–442
- **39.** Froese R, Pauly D. FishBase. In: Froese R. and Pauly D. (Editors). 2014. World Wide Web electronic publication. Available: http://www.fishbase.org.
- Palomares MLD, Pauly D. SeaLifeBase. In: Palomares MLD, Pauly D, editors. World Wide Web electronic publication. Available: http://www.sealifebase.org. 2013
- Pauly D, Christensen V, Walters C. Ecopath, Ecosim and Ecospace as tools for evaluating ecosystem impact of fisheries. ICES J Mar Sci 2000; 57: 697–706
- **42.** ICES. Report of the Workshop to draft recommendations for the assessment of Descriptor D3 (WKD3R). ICES CM 2014/ACOM 2014; 50: 1–152.
- GFCM. Report of the 13th session of the Scientific Advisory Committee. FAO Fish Aquacult Rep. 2011; 974: 1–254
- 44. GFCM. Report of the SCSA working group on stock assessment of demersal species. 2012; 1–81
- 45. GFCM. Report of the SCSA working group on stock assessment of small pelagic species. 2012; 1–258.
- **46.** GFCM. Report of the 14th session of the Scientific Advisory Committee. FAO Fish Aquacult Rep 2012; 1001: 1–200.
- 47. Froese R, Proelss A. Is a stock overfished if it is depleted by overfishing? A response to the rebuttal of Agnew et al. to Froese and Proelss "Evaluation and legal assessment of certified seafood". Mar Pol 2013; 38: 548–550.
- **48.** Rosenberg AA, Fogarty MJ, Cooper AB, Dickey-Collas M, Fulton EA, Gutiérrez NL, et al. () Developing new approaches to global stock status assessment and fishery production potential of the seas. FAO Fisher Aquac Circ 2014; 1086:1–175.
- **49.** Cullis-Suzuki S, Pauly D. Failing the high seas: A global evaluation of regional fisheries management organizations. Mar Pol 2010; 34: 1036–1042
- Carruthers TR, Walters CJ, McAllister MK. Evaluating methods that classify fisheries stock status using only fisheries catch data. Fish Res 2012; 119–120: 66–79
- Froese R, Zeller D, Kleisner K, Pauly D. What catch data can tell us about the status of global fisheries. Mar Biol 2012; 159: 1283–1292
- 52. FAO. The State of World Fisheries and Aquaculture. Rome: FAO; 2014.
- Anticamara JA, Watson R, Gelchu A, Pauly D. Global fishing effort (1950–2010): Trends, gaps, and implications. Fish Res 2011; 107:131–136
- **54.** Harley SJ, Myers RA, Dunn A. Is catch-per-unit-effort proportional to abundance? Can J Fisher Aquat Sci 2001; 58: 1760–1772
- 55. Thorson JT, Minto C, Minte-Vera CV, Kleisner KM, Longo C. A new role for effort dynamics in the theory of harvested populations and data-poor stock assessment. Can J Fisher Aquat Sci 2013; 70: 1829–1844
- Coll M, Palomera I, Tudela S, Dowd M. Food-web dynamics in the south Catalan Sea ecosystem (NW Mediterranean) for 1978–2003. Ecol Model 2008; 217: 95–116
- 57. Alheit J, Licandro P, Coombs S, Garcia A, Giraldez A, Santamaria MTG, et al. Atlantic Multi-decadal Oscillation (AMO) modulates dynamics of small pelagic fishes and ecosystem regime shifts in the eastern North and Central Atlantic. J Mar Sys 2014; 131: 21–35
- 58. Van Beveren E, Bonhommeau S, Fromentin J-M, Bigot J-L, Bourdeix J-H, Brosset P, et al. Rapid changes in growth, condition, size and age of small pelagic fish in the Mediterranean. Mar Biol 2014; 161: 1809–1822
- Fortibuoni T, Libralato S, Raicevich S, Giovanardi O, Solidoro C. Coding early naturalists' accounts into long-term fish community changes in the Adriatic Sea (1800–2000). PLOS ONE 2010; 5: e15502 doi: 10.1371/journal.pone.0015502 PMID: 21103349
- 60. Lotze HK, Coll M, Dunne JA. Historical changes in marine resources, food-web structure and ecosystem functioning in the Adriatic sea, Mediterranean. Ecosystems 2011; 14: 198–222
- 61. Stergiou KI. Fisheries impact on trophic levels: long-term trends in Hellenic waters. In: Papathanasiou E, Zenetos A, editors. State of the Hellenic marine environment. Athens: Hellenic Centre for Marine Research; 2005. pp. 326–329



- 62. Coll M, Piroddi C, Albouy C, Ben Rais Lasram F, Cheung WWL, Christensen V, et al. The Mediterranean Sea under siege: spatial overlap between marine biodiversity, cumulative threats and marine reserves. Global Ecol Biogeogr 2012; 21: 465–480
- **63.** Pennino MG, Bellido JM, Conesa D, López–Quílez A. Trophic indicators to measure the impact of fishing on an exploited ecosystem. Anim Biodiv Conserv 2011; 34: 123–131
- **64.** Shiganova TA. Invasion of the Black Sea by the ctenophore Mnemiopsis leidyi and recent changes in pelagic community structure. Fish Oceanog 1998; 7: 305–310
- **65.** Sumaila UR, Khan A, Watson R, Munro G, Zeller D, Baron N, et al. The World Trade Organization and global fisheries sustainability. Fish Res 2007; 88: 1–4
- **66.** Froese R, Proelß A. Rebuilding fish stocks no later than 2015: will Europe meet the deadline? Fish Fish 2010; 11: 194–202
- 67. Tsikliras AC. Fisheries mismanagement in the Mediterranean: a Greek tragedy. Fish Aquac J 2015; 5: 1000e113